

The Greenhouse Effect: Damages, Costs and Abatement

ROBERT U. AYRES^a and JÖRG WALTER^b

Abstract. The buildup of so-called “greenhouse gases” in the atmosphere — CO₂ in particular — appears to be having an adverse impact on the global climate. This paper briefly reviews current expectations with regard to physical and biological effects, their potential costs to society, and likely costs of abatement. For a “worst case” scenario it is impossible to assess, in economic terms, the full range of possible non-linear synergistic effects. In the “most favorable” (although not necessarily “likely”) case (of slow-paced climate change), however, it seems likely that the impacts are within the “affordable” range, at least in the industrialized countries of the world. In the “third world” the notion of affordability is of doubtful relevance, making the problem of quantitative evaluation almost impossible. We tentatively assess the lower limit of quantifiable climate-induced damages at \$30 to \$35 per ton of “CO₂ equivalent”, worldwide, with the major damages being concentrated in regions most adversely affected by sea-level rise. The non-quantifiable environmental damages are also significant and should by no means be disregarded.

The costs and benefits of (1) reducing CFC use and (2) reducing fossil fuel consumption, as a means of abatement, are considered in some detail. This strategy has remarkably high *indirect* benefits in terms of reduced air pollution damage and even direct cost savings to consumers. The indirect benefits of reduced air pollution and its associated health and environmental effects from fossil-fuel combustion in the industrialized countries range from \$20 to \$60 per ton of CO₂ eliminated. In addition, there is good evidence that modest (e.g. 25%) reductions in CO₂ emissions may be achievable by the U.S. (and, by implication, for other countries) by a combination of increased energy efficiency and restructuring that would permit simultaneous direct economic benefits (savings) to energy consumers of the order of \$50 per ton of CO₂ saved. A higher level of overall emissions reduction — possibly approaching 50% — could probably be achieved, at little or no *net* cost, by taking advantage of these savings.

We suggest the use of taxes on fossil fuel extraction (or a carbon tax) as a reasonable way of inducing the structural changes that would be required to achieve significant reduction in energy use and CO₂ emissions. To minimize the economic burden (and create a political constituency in support of the approach) we suggest the substitution of resource-based taxes in general for other types of taxes (on labor, income, real estate, or trade) that are now the main sources of government revenue. While it is conceded that it would be difficult to calculate the “optimal” tax on extractive resources, we do not think this is a necessary prerequisite to policy-making. In fact, we note that the existing tax system has never been optimized according to theoretical principles, and is far from optimal by any reasonable criteria.

Key words. Atmosphere, benefits, carbon, climate, conservation, damages, emissions, energy, greenhouse, policy.

^a Department of Engineering & Public Policy, Pittsburgh, PA, U.S.A. During the academic year 1989–90 Dr. Ayres was at the International Institute for Applied Systems Analysis (IIASA), Laxenburg, Austria.

^b Department of Computer Science, University of Illinois, Champagne-Urbana, IL, U.S.A. During the summer of 1989 Mr. Walter was a member of the Young Scientists' Summer Program at IIASA.

Glossary

B.P.	before present
CFCs	chlorofluorocarbons
CO ₂ -eq.	CO ₂ -equivalent; means RIGs, weighted in respect to radiative properties.
2 × CO ₂	doubling of sum of RIGs concentrations
GCM	Global Circulation Model
GDI	gross domestic income
GNP	gross national product
GWP	gross world product
Gt	Giga tons =billion tons
Mha	Mega ha =million hectare
RIGs	Radiatively Important Gases also called Greenhouse gases. Denotes CO ₂ , N ₂ O, CH ₄ , tropospheric ozone and CFC.
SLR	sea level rise
t-CO ₂	ton carbon dioxide per year, if used as emission or absorption rate; otherwise mentioned.

1. Introduction

Human economic and industrial activity has reached a level of intensity that threatens the stability of the global-atmosphere-biosphere system. One consequence to be expected is a significant warming of the climate. The proximate cause is a buildup of the concentrations of several trace gases in the atmosphere. The so-called Radiatively Important Gases (RIG's) are carbon dioxide (CO₂), nitrous oxide (N₂O), and methane (CH₄), tropospheric ozone (O₃), and chlorofluorocarbons (CFC's). Since pre-industrial times the first four gases have increased by 25%, 96%, 8%, and 0–25%, respectively (Ramanathan 88). CFC's are purely anthropogenic, having been invented in the early 1930's. They are used commercially as refrigerants, solvents and forming agents.

These greenhouse gases or RIG's are transparent to incoming short-wave (visible) radiation but they strongly absorb and reradiate long-wave thermal radiation. The net result is to change the radiative balance of the earth in such a way that more energy is trapped.

One of the major uncertainties in the system is ocean uptake. Today, it is generally believed that about 50% of anthropogenic CO₂ emissions are absorbed by the ocean, the remainder accumulating in the atmosphere. The accumulation (which is directly measurable, of course), results in other effects. One of them is an increased rate of photosynthesis. However, the rate of absorption (net of re-emission) by the oceans is still somewhat uncertain, and the exact role of the various actors in the system is still open to some question.

CO₂ is an essential input to photosynthesis by green plants. From laboratory experiments, it is estimated that a doubling of ambient CO₂ concentration would cause a 10–50% increase in the yield of so-called C₃ crops (e.g. wheat, rice) and a 0–10% increase in yield of C₄ crops (e.g. corn).

Depending on specific crop and growing conditions, the amount of water required to fix a unit of carbon is reduced, increasing yields in cases of growth limited by water availability (Bolin *et al.* 86). Leaf stomata, where gas exchange takes place (CO_2 in, O_2 and water vapor out) tend to decrease in size. Whether the effect of CO_2 "fertilization" will occur in open fields is uncertain. A few ambiguous multiple year experiments reported suggest no permanent increase in the photosynthetic rate (Sedjo & Solomon 90). The possibility of biochemical surprises cannot be ruled out if the concentration of a major component of organic life is doubled. (By comparison, the ambient CO_2 concentration during the last Ice Age, 18 thousand years ago, was 25% lower than it is now).

A consequence of overall climate warming is likely to be changes in the temporal and spatial distribution of temperature, precipitation, evapo-transpiration, clouds and air currents. All of these are simulated in so-called global circulation models (GCM's), although the detailed results of the simulations are not as yet a trustworthy basis for forecasting. (The next generation of such models should be considerably improved). Computations carried out to date, comparing equilibrium for the $2 \times \text{CO}_2$ condition with control runs for current climate, show very non-uniform response even to uniform change in RIGs. In effect, the regional effects are much more variable — and uncertain — than the global average projections.¹ However, the non-linear character of the system makes it likely that better GCM's will continue to exhibit significant regional variability.

The global mean temperature (GMT) is expected to rise between 2°C and 5°C for the $2 \times \text{CO}_2$ condition (Schneider & Rosenberg 90). This is remarkable compared to the last Ice Age extreme: 18 thousand years ago GMT was about 5°C colder than today (Schneider 89). The regional averages change from -3 to $+10^\circ\text{C}$ with probable changes in seasonality and variability.

Global precipitation is likely to increase by 7–10% (high confidence); regional changes are projected to range from -20% to $+20\%$ (low confidence). The largest warming will occur in high latitudes, and will be combined with large precipitation increases in winter. Higher temperatures will probably (high confidence) increase evapo-transpiration by 5–10% on global average (Schneider & Rosenberg 90). Soil moisture is controlled by precipitation, evapo-transpiration and run-off. Regional changes are projected (medium confidence) to be in the range of plus or minus 50% (*ibid.*).

Run-off would increase globally. Changes on a regional scale of -50% to $+50\%$ are expected (*ibid.*). They are direct results of changes in evapo-transpiration (which is strongly influenced by temperature) and precipitation. Simulation studies on arid and semiarid river basins in the USA suggest that relatively small changes in temperature and precipitation can have multiplier effects on run-off. There is evidence that run-off will

increase in winter in high latitudes, and decrease in summer in mid and low latitudes. These changes in run-off patterns “could greatly alter the likelihood of flooding and the availability of water during peak-demand periods such as irrigation seasons” (Frederick & Gleick 90, p. 133).

Thermal expansion of the ocean water will be the major cause of the expected sea level rise (SLR) in the short term. Robin has estimated the SLR to be in the range of 0.2 to 1.65 meters (Robin 86). Warming of the ocean is a delayed non-uniform process depending on local mixing rates. The feedback to climate will cause a transient phase which is so far not predictable with the current (equilibrium) GCM's. Impacts on ocean currents like a displacement of the Gulf Stream, or local SLR effects, are not taken account of by GCM's available to date. The major uncertainties of the current GCM's arise from inadequate knowledge of the air-ocean interface and the influence of cloud feedback. (For detailed discussion of uncertainties and model validation, see Schneider & Rosenberg 89).

2. Damages: Summary

The effects of climatic change will be superimposed on other changes, including a general increase in the intensity of land use, forest clearing, ground water withdrawal, soil erosion and air and water pollution. Acidity, of course, is a consequence of the emission of SO_x and NO_x due to fossil fuel combustion. Thus acidity and CO_2 “enrichment” of the atmosphere tend to increase together. Moreover, the environmental stresses due to acidity will tend to have a multiplier effect on the stresses of climate change. The combination will further weaken the ability of some species to survive. (The environmental acidification problem is already severe in some regions; it has been blamed for the drastic “dieback” of conifer forests in central and east-central Europe). The combination of climate change with other stresses on ecosystems could be more dangerous than any one of them taken by itself.

Impacts of sea level rise (SLR) to coastal regions are potentially massive. Coastlines will move inland up to several hundred meters, in many places, depending on beach slope and characteristics of the beach material (Hekstra 90). Salt water will also move upstream via rivers into lowland, freshwater pockets behind coastal dunes, and into ground water aquifers. The effect will be magnified in some areas where intensive ground water withdrawal has occurred (e.g. Long Island).

SLR will cause enormous loss of biologically diverse coastal lowlands and wetland ecosystems (Wilson 89). For instance, Indonesia possesses 15% of all world coastline and it is the world richest country in terms of wetland ecosystems quantity and diversity. Yet “at least 40% of its land surface is vulnerable to SLR of 1 m” (*ibid.*, p. 58). Worldwide, the land area that would be subject to inundation or made vulnerable by salt water

intrusion is about 500 million hectares. This is only about 3% of all land area, but it constitutes over 30% of the most productive cropland area (*ibid.*, p. 60). As many as one billion people now live in the vulnerable areas, including some very large cities. Thus, as has been pointed out, as much as one-fifth of world market valued assets could be adversely affected (Crosson 90).

Many environmentalists distrust economic analysis and judge the generation of quasi-market prices as ill-suited for the study of economic impacts of global climate change, for instance. They tend to advocate notions such as “safe minimum standard” as a risk-averse, conservative criterion for the survival of species, habitats and ecosystems, provided the costs are not “unacceptable large” (e.g., Batie & Shugard 90, p. 129). Yet, these “simple” policy instruments are often ineffective in practice, and they may even increase costs excessively in relation to benefits achieved.

In short, there is no real substitute for economic analysis, however unsatisfactory the present state of the subject. Nor do we distinguish (as some environmentalists do) between economic costs and “other” costs (such as eco-degradation), with the implication that the latter cannot be compared with (or traded off against) the former. To us, it is a question of defining the realm of economics broadly enough. All costs are economic if the economy is properly defined, but not all economic costs are automatically reflected in the marketplace (i.e. “market valued”). Nor are non-market-priced environmental assets (such as parks or ecosystems) included in the standard System of National Accounts (SNA).

3. Nordhaus' Estimates of Damage Costs

A recent study by William Nordhaus has attempted to estimate the economic costs of climate change (Nordhaus 89, 90). He began with a breakdown of the U.S. gross national income or GNI² (for 1981) by sector and subdivided it further into “regimes” of sensitivity. The most climate sensitive sectors were agriculture, forestry and fisheries, which amounted to 3.1% of total NI. Moderate sensitivity was attributed to sectors such as construction, water transport, utilities etc. These contributed 10.1% of the total. The rest (86%) comes from sectors affected negligibly by climate (e.g., mining, finance, manufacturing etc.)

The results of this analysis were as follows:

1. Agriculture damage costs (offset by the CO₂ fertilization effect) are estimated as plus or minus \$10 billion as an overall impact on all crops.
2. Sea level rise (SLR) damages were estimated for land loss (15,540 square km) and protection of high value property and open coasts by levees and dikes. The total market value of the property at risk is on

the order of \$100 billion. Nordhaus converted this to an estimated annual equivalent loss of \$6.18 billion per year. (The capital value of property should reflect its continuing flow of benefits, thus reflecting tourism losses implicitly, at least so far as providers — hotel and motel operators and so on — are concerned. What is omitted is the loss of use and option value to users, who may not be able to find equivalent amenities elsewhere).

3. Greenhouse warming is expected to increase aggregate demand for air-conditioning (\$1.65 billion/yr) and reduce the demand for space heating (\$1.16 billion/yr). Assuming average current prices for electricity and fuel, there would be a net annual extra cost to the economy of \$0.46 billion (USEPA 88).
4. No specific estimates were made by Nordhaus for other goods or services (either market-valued or otherwise). In effect, these were lumped together and included in the uncertainty of the overall estimate (see below).

Summarizing the quantified cost items above, the breakdown is as follows:

8% attributable to energy demand changes
 92% attributable to SLR (of which 85% is for coastal protection cost — levees, seawalls, etc. — and 7% is for loss of low-lying land)

The “bottom line” — the central (most likely) estimate of total annual economic damages — was \$6.67 billion (1981 dollars), assuming the damages occurred in 1981. This is equal to 0.28% of U.S. gross national income for that year. The error bounds were judged (by Nordhaus) to be quite a bit higher, due to the omitted unquantified items, but still less than 2 percent of national income.

Gross world income (GWI) in the year 2050 is likely to be more than \$26 trillion 1981 dollars (USEPA 89, low GNP-case). This is 8.1 times more than U.S. national income in 1981. Thus Nordhaus judged this scaling factor of 8.1 to be appropriate to extrapolate the US damage “snapshot” to global annual damages in 2050 (assuming similarity of income structures). In other words, annual world damages due to greenhouse warming are “most likely” to be about \$54 billion (\$1981), with an upper limit of \$520 billion. Based on expected emissions of 16.9 billion tons of CO₂ equivalents Nordhaus converted this to *marginal shadow damages of emission*, viz:

central case:	\$3.3/ton (CO ₂ -equiv.)
worst case:	\$36.9/ton (CO ₂ -equiv.)

The above calculation (Nordhaus’ numbers) is based on one fairly “heroic” (and technically incorrect) assumption with regard to physical

damage: that future damage is simply proportional to RIG emission rates on a *current* basis, i.e., no accumulation of RIGs, and no damage dependence on rates of warming. Of course, the economic assumptions are equally strong, as already noted. For example, the pattern of energy use in the U.S. bears little relation to the rest of the world. The extrapolation to global scale assumes a similar balance between air-conditioning and space heating, which is somewhat implausible.

4. Modifications to Nordhaus' Estimate

Bearing in mind the long list of potential adverse effects and costs, most of which have not been quantified — or even mentioned — by Nordhaus, many environmentalists will not be satisfied with the relatively simplistic sort of calculation exemplified above. To address these doubts it seems useful to examine Nordhaus' assumptions in more detail. We focus, first, on the implications of sea level rise (SLR), inasmuch as this item accounts for 92% of the total costs identified by Nordhaus.

With regard to SLR the major costs identified above are protection costs of valuable coastal land and beaches (via seawalls, dikes, and levees). The total U.S. coast length is about 20,000 km. Average protection costs of about \$5 million per km coastline appears reasonable in view of Dutch experience (e.g., Hekstra 90).

The coastline of the world amounts to between 0.5 million and 1 million km. To protect it to the same extent as projected in the US, the total cost would be about \$2.5–\$5 trillion, or 10–20% of minimum GWI for 2050. Spread in proportion to GWI over 50 years, as Nordhaus did, this comes to about 0.2–0.4% of world GWI annually, or roughly what Nordhaus assumed. It is a rough magnitude of avoidance costs for physical protection of “protectable” low lying areas, estuaries and so on.

Nordhaus' estimate of land-loss cost of 1.55 Mha (million hectares) along the U.S. coastline (19,924 km) is equivalent to 77 ha/km coastline. This is a factor of ten less than Hekstra's estimate of 500 Mha vulnerable land along 0.5–1.0 million km coastline, or 500–1000 ha/km (Hekstra 90). The land value assumed by Nordhaus (\$5000/ha) lies in between Hekstra's estimate for arable cropland in Bangladesh (\$3000/ha) and in the Netherlands (\$30,000/ha). Assuming Nordhaus' price of \$5000/ha the total land value loss based on Hekstra's estimate of vulnerability, would be \$2.5 trillion. Spread over 50 years this would account for 0.3% of the world GWI, on average. This is still well within Nordhaus' range of error, of course.

Yet the methodology of estimating potential loss by attaching current values to submerged land is inherently suspect, even allowing for “scaling”. In the first place, current monetary prices of land in different countries clearly reflect current levels of money income and exchange rates. In the

second place, since the total amount of arable land will be reduced in absolute terms, it is clear that the price of the remaining land will rise along with the sea level. The gain in land values elsewhere could well outweigh the coastal losses. Yet one could hardly conclude that SLR might therefore be beneficial. Moreover, the remaining land would have to be cultivated more intensively to make up the shortfall, and food prices will rise, as Schelling noted. A gain for the (remaining) farmers, but a loss for consumers. (The same valuation problem arises if OPEC succeeds in raising the price of oil).

The use of land prices (based on current exchange rates) implies that coastal land in the U.S. or the Netherlands is more valuable than coastal land in Bangladesh or the Nile Delta. This conclusion makes no sense for a study of this kind. Land is more productive in Bangladesh or the Nile Delta than in the U.S. and probably no less productive than in the Netherlands. Land value should be related to its productivity in *real* terms for purposes of assessing long term costs of climate warming. On this basis, land losses in Bangladesh or Egypt should be evaluated at \$30,000/ha, rather than \$3000/ha. Using prices based on international exchange rates undervalues land in poor countries by an order of magnitude. Moreover (as Nordhaus noted) the U.S. derives little of its national income from coastal lands; the opposite is true in Bangladesh. A loss of 10% of the arable land of a country where 70% of the population lives on the land would (roughly) cut its real national income by at least 7%. It is the exchange rate that is artificial and misleading (being based on trade balances in a few portable commodities and manufactured goods). If the notion of marginal utility — rather than land price — were invoked, it would seem to follow that the utility loss to Bangladesh must be far greater, per capita, than the utility loss to the U.S. Thus, the extrapolation from U.S. calculations to the third world is unsatisfactory, to say the least.

Since the vulnerable low-lying lands are heavily populated, we must expect some environmental refugees. For example, more than 1000 islands in the Maldivian Atolls may be swallowed up by the sea. The deltas of the Brahmaputra River (Bangladesh) and the Nile River (Egypt) are densely populated. Assuming SLR of 0.79 m by the year 2050 and 2.17 m. by 2100, the homes and livelihoods of 46 million self-supporting people would be lost (Jacobson 89). Under “really worst case” assumptions, including widespread subsidence due to excessive groundwater pumping, the number threatened would be substantially higher. Bearing in mind Hekstra’s estimate of one billion people potentially “affected” by SLR, it is not unreasonable to suppose that as many as 100 million people — mainly subsistence farmers with no urban experience or skills — may be displaced. They will have no place to go except to the already overcrowded cities.

How much does a refugee cost? It depends where the refugees are located and on their status and skills. Malawi’s social cost per Mozam-

biquan refugee is reported as a mere \$24 per capita (The Economist, February 18, 1989). An inquiry by the UN High Commissioner of Refugees and the World Food Program sets the annual average expenditure of these two official institutions per assisted refugee at \$72 per capita, or about 20 cents per day; not too much. These costs reflect extremely bad conditions, such as those in camps for Palestinian refugees located in Lebanon and Jordan. On the other hand, the U.S. spends some \$4000 per accepted refugee (\$362 million for 94,000 refugees arriving in 1988 (The Economist, September 24, 1988)).

These are just maintenance or resettlement costs. Since a refugee is obviously unproductive for some time, it would be sensible to assume one or more years of lost output (GNP/capita). In the case of the “low cost” Palestinian refugees, there is no resettlement program and the production loss is much more than a year or two — more nearly permanent. The social costs of repression, terrorism, regional political turmoil, and military/police responses to all of the above should be included also. These costs tend to dwarf the pure “subsistence” costs, although they are almost never properly allocated. Even in the case of refugees admitted to the U.S. or other industrialized countries, the period of adjustment is significant, especially if the refugees are uneducated. In order to get a crude magnitude of likely social costs for resettling economic refugees from the poorer countries (within the same country) we assume a modest two year period of lost output at \$250/yr, or \$500/capita at 1981 income levels. (Comparable GNP/capita figures for 1985 were: India \$270, Bangladesh \$160, Egypt \$760 (WRI 89, p. 236). Altogether, this adds up to \$250 billion, over 50 years. Assuming significant economic growth in these areas, resettlement costs and losses rise in proportion; it would not be unreasonable to double or even quadruple this figure. To be conservative, we doubt it.

A revised set of SLR costs, based on the above reasoning, is as follows:

Coastal protection cost: \$2.5–5 trillion

Coastal land loss: \$15 trillion

Costs of resettling 100 million refugees @ \$1000 each: \$1.0 trillion

Total: \$18.5–21 trillion.

Of course, this is a total for the world as a whole, spread over 50 years, as Nordhaus did, and therefore comparable to his numbers. Annualized, it comes to around 2.1–2.4% gross world income (GWI), or nearly 10 times higher than Nordhaus “central” estimate for total costs, and slightly outside his range of error. For purposes of analysis, therefore, we think \$30–35 per tonne of CO₂ (equivalent) is more realistic than \$3.30, just to take account of the effects of SLR on countries like Egypt and Bangladesh.

To be sure, many indirect effects are still omitted, which have completely unknown shadow costs. One of the most obvious is the implicit

assumption that there is empty land available somewhere to resettle the refugees. In fact, there is no likelihood of such resettlement. Displaced persons will crowd into cities creating squatter settlements that tax the available city services to the limit. These shanty towns are already prime reservoirs of frustration and disaffection, and a breeding ground for violence, crime and civil unrest. What are the true social costs of uprooting people, taking into account the breakdown of traditions and family relations, and the resulting social problems for the rest of society? We do not know, except that the costs are not zero.

Moreover, large numbers of refugees in Southeast Asia would augment the immigration pressure to the more highly developed countries in a dramatic way. The "boat people" from Viet Nam may be only the vanguard of an enormous migratory wave the world in general (and Australia, in particular) is ill-prepared to cope with. So far, the USA hasn't succeeded in integrating its black population, after 125 years of struggle. Britain hasn't solved its problem with the commonwealth immigrants, France has difficulties with the North African immigrants, while West Germany is finding its small Turkish minority quite indigestible. Lacking adequate "social technologies" most countries will end up spending more money, instead, on internal and external security.

In summary, there is good reason to believe that "when the winners and losers have been identified, there will be little interest on the part of the winners to alter their status in order to compensate the losers" (Glantz 88, p. 409). In short, there is increasing risk of tensions, frictions and conflicts threatening to political stability.³ Yet, it is impossible to put a convincing number on these indirect effects, if only because the causes of social tensions and disruptions are multifarious and the "greenhouse effect" contribution is likely to be relatively minor compared to other factors. All things considered, Nordhaus' estimates seem too optimistic by a considerable margin.

Before moving on to consider abatement strategies and costs, it must be pointed out once again that Nordhaus' estimates of losses and costs exclude all losses to final users of environmental assets, as well as option and bequest value losses. What is the option value of the last Redwood forest or the large shade trees on urban streets and in urban parks? Old, slow-growing trees like oaks, elms, maples and beeches are clearly vulnerable to climate change (cf. the work of Leemans and Solomon, cited earlier) and are highly valued. Since fully grown trees cannot be moved, there is no actual market for them; however, the retail prices of relatively young trees (around 20 years old) range up to \$500. It is quite normal for suburban property owners in the U.S. to spend several hundred dollars per year for tree care.

If this can be taken as an indicator of the value of the underlying assets, then one would have to impute a value of at least several thousand dollars

to each mature shade tree in a built up area. The number of such trees is unknown, but it probably exceeds the number of people (at least in the U.S. and Western Europe), If the life expectancy of shade trees is reduced from 200 years to 50 years by rapid climate change, there will be a major loss of amenity value, and a sharp increase in expenditure on landscaping (the rate of tree-planting would have to increase by 4-fold, for instance). Other costs of maintaining parks and gardens will also rise sharply. This would translate into significant annual costs for both individual homeowners and cities. We don't attempt to take the calculation further, except to note that annual expenditures by suburban homeowners of the order of 2% or 3% of income to maintain trees and shrubs are by no means uncommon today. (Averages are smaller, of course). Still, an annual average expenditure for this purpose in the next half century (including indirect outlays) attributable to the higher costs of compensating for effects of climate change, would not be implausible.

In summary, we suspect that the sum total of potential losses of this type greatly exceeds the items that Nordhaus has actually quantified.

5. Optimal Abatement

The usual hypothetical relationship between emissions damages and abatement costs (Figure 1). Assume we know all damages $D(z)$ as a function of annual emission z of greenhouse gases, incorporating present and future values, priced and unpriced (see discussion above). Further, assume we know the cost function of abatement $A(z)$ for all levels of emissions. By assumption $A(z)$ describes the total cost to an economy to abate the next

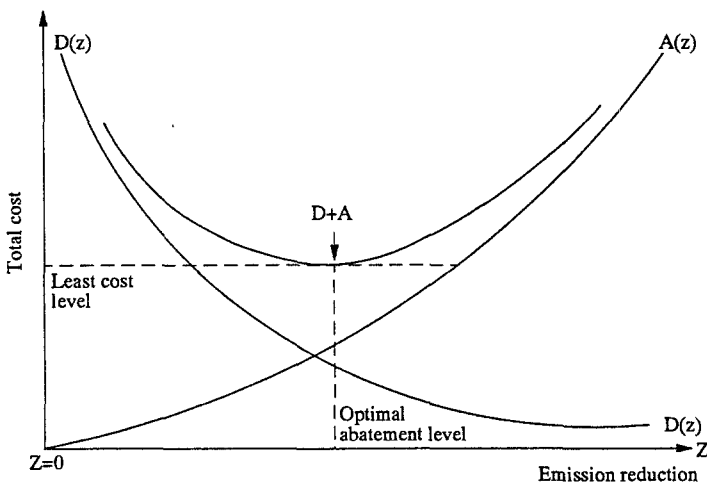


Fig. 1. Optimal abatement in economic equilibrium.

increment of greenhouse emissions by the most cost-effective available means. If reducing fossil fuel combustion is the chosen strategy, then the cost curve would reflect the costs of introducing energy-conserving technologies or providing alternative fuels, for instance.

The shape of $A(z)$ is usually derived from two general axioms in economic theory, namely, (i) that the economy is always in or (nearly in) an equilibrium state and (ii) declining marginal cost-effectiveness of abatement with increasing levels of abatement. Given these assumptions, abatement costs are zero at the "Laissez-faire" point (of uncontrolled emissions) and increase as a function of increasing abatement (CO_2 -equivalent reduction). Because of declining marginal cost-effectiveness, the real cost of abatement $A(z)$ can be expected to increase *at an increasing rate*, as shown.

The optimal level of abatement is the minimum of the sum $D(z) + A(z)$, which has, by definition, a slope of zero. Both less and more reduction would lead to reduced welfare benefit. In other words: the optimal point is characterized by equality of the absolute first derivatives (marginal costs) of A and D . Evidently, the *marginal benefit* of abatement is obtained from the slope of the damage cost curve, $D(z)$. The marginal cost of abatement is the slope of $A(z)$, where z is measured in percentage of CO_2 -equivalent reduction.

Figure 2 contrasts with the usual version (Figure 1) with a rather different form of the abatement cost curve. It is inconsistent with one of the two key assumptions underlying Figure 1 (the equilibrium assumption), but we think it comes much closer to reflecting reality. It reflects the view that,

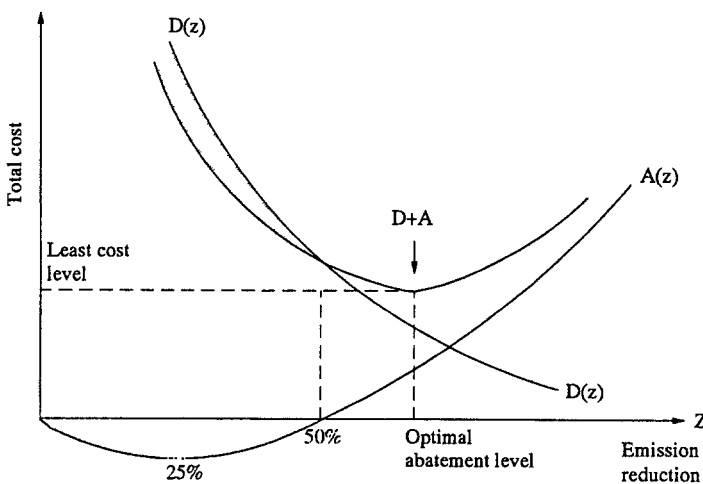


Fig. 2. Optimal abatement in economic disequilibrium.

in fact, there exists a considerable opportunity to enjoy *negative abatement costs* (i.e. profits) by investing in selected technological “fixes”, largely in the area of energy conservation.

The implication, of course, is that, for a variety of reasons — including massive market failures — the economy has become “locked in” to a sub-optimal state of excessive energy/resource dependence.⁴ We defend this proposition in more detail below. Of course, the optimum degree of abatement is still the point where $D(z) + A(z)$ is minimal. However, it will be noted that the optimal point is significantly to the right of the corresponding point in the case where $A(z)$ is monotonically increasing (Figure 1) and — more important — the optimal abatement level is far greater.

In the following discussion we focus mainly on the costs and benefits of CO₂ emission reduction by energy conservation and (to the extent feasible) fuel substitution. Nordhaus also considered two other possibilities: reforestation and CFC reduction. We review the latter two options briefly in Appendices A and B respectively.

6. Costs of Combustion-Related Emission Reduction

Ranking various alternative sources of energy and possibilities for switching to less carbon rich fuels in terms of cost-effectiveness enables one to construct an abatement cost curve such as $A(Z)$ in Figures 5 or 6. Assuming the energy supply/conversion and industrial component of the economy “optimizes” quickly to adjust to changing prices (hence, it is always in or near its instantaneous equilibrium) such models can be used to estimate cost curves for various policy assumptions. Nordhaus has, for example, estimated the costs of achieving a given energy output with successively lower amounts of CO₂ production (Nordhaus 73, 75). He found that shadow costs could be expressed by a quadratic function of percentage emission reduction. A similar result was later obtained econometrically (Nordhaus & Yohe 83). Nordhaus used this function (updated to 1989 prices) in his recent work (Nordhaus 89, 90).

Another instance of this ‘macro’ approach is found in the work of Manne, Richels and Weyant.⁵ This is a major modelling effort linking a macro-economic model and an energy supply-conversion optimization model of the “activity” type. There are three underlying assumptions: (1) that the economy is always in a quasi-equilibrium state, (2) that it “finds” the optimum supply mix for a given demand more or less instantaneously, and (3) that energy consumption is both an input (factor) and a cost of production and a claim on resources. The former assumption means that energy appears as an input in a production function. When the production function is econometrically fitted to past data on energy consumption, energy prices, and total output of goods and services, it is possible to estimate the reduction in output. This can be interpreted (somewhat

loosely) as the economic “cost” of reducing energy inputs by a given amount.

The interpretation of ‘lost gross output’ as ‘cost of change’ is justified, for most economists, by the notion that GNP is a measure of aggregate social welfare. This interpretation has been criticized, for various reasons. However, we do not propose to review the arguments *pro* and *con* further here.

Engineers and businessmen think of costs in a somewhat different and more traditional way. A businessman would try to compute cost as the annualized net *additional* capital and operating costs of investing in and using a new technology. It can happen, of course, that little or no new investment is needed or that the result of the substitution results in a net saving, rather than a net cost.

For a business or a householder, a “net saving” translates into a profit, or a return on investment. The usual standard of comparison is money invested in high quality government bonds or, simply “money in the bank”. In other words, if a given investment produces a greater return (assuming equal risk) than money invested at the current rate of interest, it is “profitable” in the above sense. If the rate of return is less than the interest rate, the investment is a loser. The usual target rate of return-on-investment (ROI) for business investments — which tend to be fairly risky, and which must allow for taxes on the profits — is typically around 30% per annum. If the best return that can be realistically expected is only 15%, a prudent businessman will not make the investment. On the other hand, for a government (which does not have to pay taxes and can borrow money at lower rates than a private business); an 8% or 10% expected rate of return is probably adequate justification.⁶ (This is often equated roughly with the social discount rate).

Given that capital is scarce, it is rational to invest in the most profitable ventures first. Thus, a business will typically try to rank order the various proposals for capital spending (in order of expected ROI) and go down the list until either the available money for investment runs out or the threshold is reached. In principle, government would do the same. In a quasi-equilibrium economy, there should be enough capital to fund all of the promising projects, i.e. all the projects with expected ROI above the appropriate threshold level. It follows that the really “good” (i.e. profitable) projects should be funded as soon as they appear on the horizon. In an economy very close to instantaneous equilibrium there should be very few investment opportunities capable of yielding returns far above the average. (In fact, the average return after taxes should be the same as the rate of GNP growth). The existence of many opportunities is an indication of significant deviation from instantaneous equilibrium.

In this context, it is relevant to note that most large-scale energy *supply* projects (e.g. hydroelectric or steam-electric plants) yield a long-term real net rate of return between 5% and 10% (Economist, January 6 1990, p.

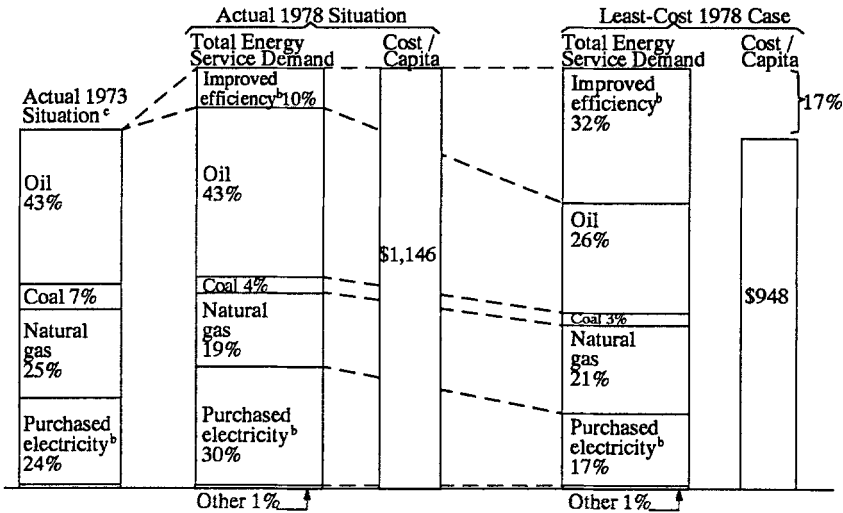
59). Since this is below the threshold level for a rational tax-paying profit maximizer, it is difficult not to suspect that non-economic factors are involved in diverting capital into such investments.

On the other hand, there is ample evidence of under-utilization of profitable opportunities for conserving energy. In a major study carried out by the Italian energy research institute ENEA it was shown that technological “fixes” exist with payback times of 1–3 years — well below the typical threshold for most firms and several times faster than investments in new supplies (e.g., d’Errico *et al.* 84).

Even more convincing evidence comes from the experience of the Louisiana Division of Dow Chemical Co. in the U.S. In 1981 an “energy contest” was initiated, with a simple objective: to identify capital projects costing less than \$200,000 with payback times of less than 1 year (Nelson 89). In its first full year (1982), 38 projects were submitted, of which 27 were selected for funding. Total investment was \$1.7 million and the 27 projects yielded an average ROI of 173%. (That is, the payback time was only about 7 months). Since 1982, the contest has continued, with an increased number of projects funded each year. The ROI cutoff was reduced year-by-year to 30% in 1987, and the maximum capital investment was gradually increased. Nevertheless, in the year 1988 95 projects were funded, for a total capital outlay of \$21.9 million and — surprisingly — an average ROI of 190%! The average submitted ROI for 167 audited projects over the entire 7 years was 189%, while the actual (post-audit) average was 198%. Table I summarizes the results of the Dow experience.

It is important to note that, although the number of funded projects increased each year, there is (through 1988) no evidence of saturation. Numerous profitable opportunities for saving energy, with payback times well below one year, apparently still exist at Dow even after the program has been in existence for 7 years. One would have to suspect that the program could still be expanded many-fold before reaching the 30% ROI threshold. Furthermore, it is important to emphasize that these opportunities exist even at relatively low U.S. energy prices. Should taxes or a new energy crisis force U.S. prices higher (i.e. toward world levels), the number of such opportunities would be multiplied further.

At the macro-level, it has been argued by the Mellon Institute that a “least cost” strategy for providing energy services for the U.S. in 1978 would have utilized much less primary energy, and in a very different manner, than that which was actually observed (Sant 79). In economic terms, the least-cost strategy would have saved \$800 per family (17%) or \$43 billion in that year alone (*ibid.*). Taking the year 1973 as a standard for comparison, such a strategy would have involved a sharp reduction in the use of centrally generated electricity (from 30% to 17%) and a reduction in petroleum use from 36% to 26%. The only primary fuel to increase its share would have been natural gas (from 17% to 19%). Interestingly, the



^a The primary fuel equivalent of service demand in 1978 was 79.0 quads, plus 9.2 quads of improved efficiency (calculated against a base of stock & equipment in place in 1973) or a total of 88.2 quads. Actual service depends on the conversion efficiency of fuels & equipment utilized.

1 quad = a quadrillion = 10¹⁵ BTU. Another means of visualizing a quad is 1 million barrels/day of oil equivalent = 10¹⁵ BTU (quads)/year.

^b In terms of primary fuel. ^c Primary fuels demand in 1973 was 74.6 quads.

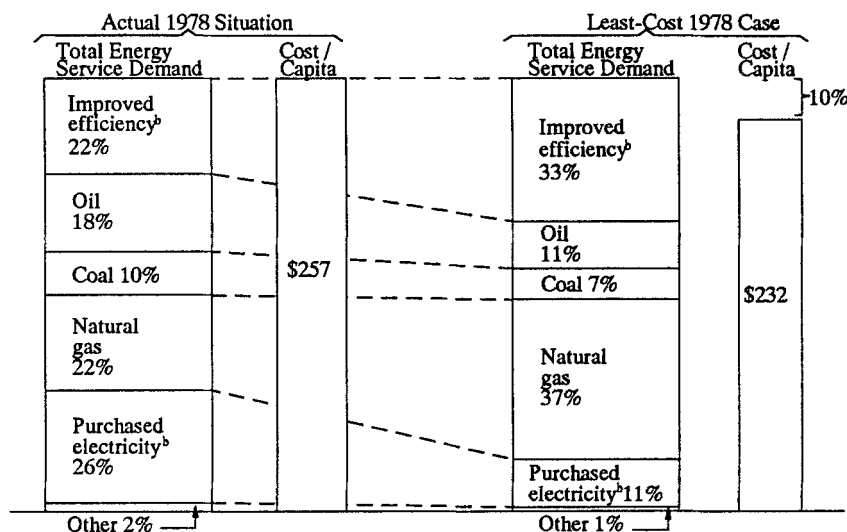
Fig. 3. Energy sector market shares^a of various technologies. Source: (Sant 79).

Table I. Summary of Louisiana division contest results — all projects

	1982	1983	1984	1985	1986	1987	1988
Winning Projects	27	32	38	59	60	92	95
Capital, \$MM	1.7	2.2	4.0	7.1	7.1	21.8	21.9
Average ROI (%)	173	340	208	124	106	77	190
ROI Cut-Off (%)	100	100	100	50	40	30	30
Savings, \$M/yr							
Fuel Gas ^a	83	-63	1506	2498	798	2550	10790
Capacity						1197	2578
Maintenance	10	45	-59	187	357	2206	583
Miscellaneous						19	-98
Total Savings	1590	3838	5341	7353	6894	11944	18023

^a All fuel gas savings are based on 1988 incremental fuel gas value.

Source: (Nelson 89), Table 1.



^a The primary fuel equivalent of service demand in 1978 was 28.2 quads, plus 7.9 quads of improved efficiency & 0.8 quads of biomass (calculated against a base of stock & equipment in place in 1973) or a total of 88.2 quads.

^b In terms of primary fuel.

Fig. 4. Industry energy service market shares^a of various technologies. Source: (Sant 79).

Mellon study suggested that “conservation services” would have increased their share from 10% to 32% in the optimal case. See Figures 3 and 4.

What the Mellon study showed, in fact, is that (at 1978 energy prices) conservation, up to a point, would not have cost *more* (as Figure 1 implies) — it would have cost *less* (as in Figure 2).⁷ Between 1973 and 1978 “conservation services” reduced actual energy consumption by 10% compared to the 1973 baseline pattern; but a 32% reduction would have been not only possible, but cheaper! The differential between actual and potential is 22%. The greatest potential savings were to be found in the so called “buildings” sector (23%), but non-trivial savings (10%) were also available in the industrial sector. Significant opportunities also existed in industry, notably by avoiding (or in other words, using) waste heat by means of heat-cascading, heat pumps and co-generation (of electricity and process heat).

Assuming the Mellon Institute’s figures to be roughly correct, the 22% unachieved but possible energy savings in 1978 would have reduced carbon dioxide production by at least 25% as compared to actual emissions. This amounted to around 275 million tons. The monetary savings to energy consumers would have been \$43 billion, as noted, or about \$65 per ton of CO₂ saved. This point would define the low point of the (negative) marginal cost of abatement curve in Figure 2.

If the curve in Figure 2 is *symmetric* on both sides of this point, it follows that a gross reduction in fossil energy use and CO₂ emissions of roughly 50% should have been achievable (in 1978) at *zero* net cost to the economy as compared with the actuality. Clearly the optimum abatement level would be somewhat further to the right, perhaps around the 60% level, depending on the value of reducing emissions.

Admittedly the 1978 disequilibrium might have been reduced somewhat in the last decade, and energy prices have (temporarily) dropped. The above calculation is illustrative, at best. Still, an increase in energy prices (via taxes) would merely increase the already clear benefits of investing in energy conservation. Others have arrived at similar conclusions (see, for example, Lovins *et al.* 81). There is growing evidence supported by numerous examples that many investments in energy conservation can pay for themselves in reduced operating costs in a few months to a few years, even at present (lower) energy prices.

Before moving on to other issues, it is worthwhile commenting on the evident discrepancy between actual behavior and optimal behavior. Consumer behavior with regard to energy conservation is sometimes dismissed as “stupidity-limited” (Schipper 89). However, what appears to be “stupidity” at first glance can be resolved into two other phenomena. One of the probable explanatory factors is inadequate information. (Information is not costless, in the real world). It has been remarked that lack of awareness by the consumer is a major impediment to increasing conservation in the end-use sector. Consumers focus today mainly on product color, size, and features and only a small amount on energy consumption during the appliance’s lifetime. To some extent this can be countered by public information and awareness campaigns and government-sponsored information programs, such as “green labelling”.

The other phenomenon, which deserves more study, is that individuals seem to display extremely high *discount rates* in their personal financial affairs. To put it simply, people will not voluntarily pay much extra or wait very long for promised savings in future operating costs of houses, automobiles or appliances. One study, based on detailed survey data on household appliance purchases, has inferred the average discount rate for consumers to be 20% (Hausman 79). It is strongly income dependent, however, as shown in Table II. Note that much higher discount rates are observed for households with very low incomes (89% for the poorest category). Translated into payback times, consumer behavior among the lowest income group in the U.S. seems to correspond to personal “payback times” of the order of 1 year (or even less).

Any investment that pays for itself in much less than twenty years would be an unambiguous economic gain at the macro-economic level. (This follows from the fact that the average growth rate of the economy corresponds in a doubling time of the order of 20 years). Businesses are

Table II. Estimated discount rates using mean population estimates

Income Class	# of Observations	Implied Discount Rate
\$6,000	6	89.0%
\$10,000	15	39.0%
\$15,000	16	27.0%
\$25,000	17	17.0%
\$35,000	8	8.9%
\$50,000	3	5.1%

generally happy to invest in moderately risky projects that will pay for themselves in 5 years if they are successful. The existence of numerous opportunities for paybacks of one or two years, with almost no risk, is clear evidence of a non-equilibrium situation.

Many economists have trouble with these implications. They ask: if, indeed, such opportunities really exist, why don't entrepreneurs operating in the competitive market-place find and exploit the opportunities? The fact that this doesn't seem to be happening suggests that the opportunities are not real, after all; according to this view, there must be "hidden costs".

We have no authoritative answer to the question. (It is a topic that should be given greater attention by business schools, among others). We think, however, that the basic answer is related to two facts: (1) the opportunities for energy conservation are mostly incremental; they require many small investments, rather than a few massive ones. This is difficult for industry because of the second fact (2) that large firms are central planners; they do not operate *internally* like competitive markets. They are bureaucratic, hierarchical and "rule driven" rather than competitive. People with entrepreneurial instincts generally find it very difficult to function in bureaucracies or large firms. By the same token, large firms find it very difficult to induce their employees to behave entrepreneurially. The reasons for this must be sought in the incentive systems that function in bureaucracies.

7. Secondary Economic Benefits of CO₂ Reduction

As noted above (Figures 1 or 2), the economic benefit of emission reduction is equated with the corresponding damage reduction. A simple approximation to damage reduction in the present case can therefore be obtained by dividing total "greenhouse" damages by total carbon dioxide. It was suggested by Nordhaus that the most probable cost of such damages would be \$3.30 per ton of CO₂ eliminated, with an upper limit of \$37/ton (Nordhaus 89). We argued above that the first figure may be too low by a

factor of ten, not even allowing for unquantified items such as loss of recreational and ecological assets. For purposes of further analysis we take the direct "Greenhouse" benefits of eliminating CO₂ to be at least \$30 per ton, as discussed previously.

However, except for the deforestation component, carbon dioxide is almost entirely produced by the burning of fossil fuels. This, in turn, generates emissions of air pollutants such as SO_x, NO_x, CO, and so on. One cannot eliminate combustion-related CO₂ without cutting down on the other pollutants, barring the unlikely case that all of the reduction is accomplished at the expense of the most benign of the fossil fuels, natural gas. In economic terms, greenhouse gases and conventional air pollutants are *co-products*. Hence the benefits of CO₂ emission reduction must be equated to the full set of benefits of reduced fossil fuel (at least, coal and oil) combustion, whether it be achieved by regulation, conservation or taxation. The latter, in turn, include all of the health and environmental benefits of air pollution reduction. (It is important to observe that air pollution benefits due to reduced fuel consumption will be in addition to and independent of any benefits achieved by emission controls).

Of course, the air pollution and health costs of fossil fuel use depend on the specific fuel. They are much greater for coal, for instance, than for natural gas. One can monetarize these benefits by a procedure that was first used for evaluating the social costs of road traffic and energy consumption in West Germany (Grupp 86; Hohmeyer 88). The major air pollutants cause damages at very different concentration levels. Relative weighting factors can be chosen as follows: Particulates = 100; SO₂ = 100; NO_x = 125; VOC = 100; CO = 1.

Since these toxicity weightings are derived mainly from animal experiments, the extrapolation to impacts on vegetation and materials remains questionable. However we know that SO₂ and NO_x contribute roughly equally to acidification, while CO, NO_x and the hydrocarbon components take part in photochemical reactions (leading to ozone production) in a rather complicated and interrelated manner. Since most of the monetarized damages are linked to human health, these assumptions seem to be a defensible compromise.

We also assume that the flow of annual damages from pollution (exclusive of the greenhouse effect) is approximately proportional to current emissions. This means no accumulation effects are considered. Since the scope of our concern is global, no complex trans-border transport of pollutants need be considered.

Table III shows emissions of traditional air pollutants in the FRG and the USA in the years 1975–1985, from all sources (including mobile sources (MS) and power plants (PP)). These two categories account for most of the coal and petroleum, although some gas is used in electricity production and some oil is used for home heating. In effect, we are

Table III. Air pollutant emissions fraction by sector (1000 t/yr)

	FRG			USA		
	1975	1980	1985 ^a	1975	1980	1985 ^a
Particulates						
all	813	696	576 ^a	10600	8500	7000 ^a
MS	61	64	70 ^a	1300	1300	1400 ^a
PP	179	151	127 ^d	3036	2125	1750 ^b
SO₂						
all	3325	3187	2345 ^a	26000	23900	21600 ^a
MS	132	107	94 ^a	700	900	900 ^a
PP	2062	1976	1426 ^d	16829	16102	14768 ^c
NO_x						
all	2532	2935	2924 ^a	19100	20300	19800 ^a
MS	1297	1594	1718 ^a	8900	9200	8800 ^a
PP	709	813	719 ^d	4929	6101	6463 ^c
HC						
all	2545	2486	2371 ^a	22800	2300	20300 ^a
MS	1164	1249	1196 ^a	10200	8200	6700 ^a
PP	25	16	24 ^d	222	230	203 ^b
CO						
all	13014	11708	8804 ^a	81000	76100	64300 ^a
MS	10148	8808	6301 ^a	62000	52600	45200 ^a
PP	521	471	352 ^a	3380	3044	2572 ^b

^a (OECD 89).^b adjusted from (Benkovitz 82).^c adjusted from (NAPAP 90).^d adjusted from (Hohmeyer 88).

neglecting the pollutant effects of natural gas consumption. Weighting the data with the (toxicity) factors given above yields the results in Figure 5. Table IV shows the result in relative shares for mobile sources and for the power plant sector in total air pollution for the FRG and USA, Table V gives total annual CO₂ emission in these two sectors for each country.

Table IV. Air pollution fractions (% of total) from weighted emissions by sector

	1975	1980	1985
F.R.G. Mobile Sources	30.9	34.5	39.5
Power Plants	31.6	31.2	27.4
U.S.A. Mobile Sources	28.5	27.5	27.5
Power Plants	31.3	32.0	33.4

Sources: (OECD 89; NAPAP 90(II); Hohmeyer 88).

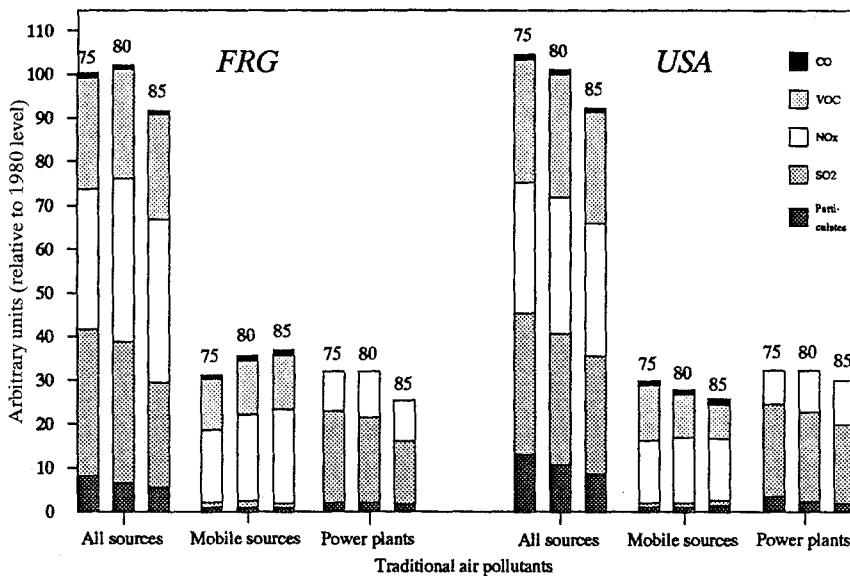


Fig. 5. F.R.G. & U.S.A.: weighted emissions 1975–1985. *Source:* authors (adapted from OECD 89 & others).

The highest costs of external effects of air pollution are health-related.⁸ Respiratory diseases, for instance, lead to costs of medical treatment, increased morbidity (loss of working time plus direct costs of illness) and an increased risk of mortality. Many epidemiological studies have attempted to establish a relationship between ambient pollutant concentration and incidence of acute and chronic diseases. Based on such studies (especially Lave & Seskin 71), Myrick Freeman estimated a pollution-mortality elasticity of 0.05, which means a 0.5% decrease in mortality for a 10% pollution reduction (Freeman 79, 82). He concluded that a 20% air pollution reduction in the U.S. would save between 2780 and 27,800 lives per year. (Freeman's work is old, but has not been superseded).

The next step is to impute the monetary valuation of morbidity and mortality benefits. While this is a highly controversial topic, it is unavoidable. In many policy areas it is necessary to balance the value of saving human life (by investing in risk-reduction) against other benefits, implicitly if not explicitly. Freeman used a value of \$1 million per statistical life saved, based on individuals' apparent willingness to pay to reduce risk of mortality (Freeman 79). Others have used the loss of expected net future domestic income per capita to estimate the damages to society which range from \$0.3 to \$10 million per human life (e.g. Gaines *et al.* 79; Grupp 86; Wicke 86).

Freeman's best point estimate for the annual human health benefits of a

Table V. Energy consumption & related CO₂ emissions — 1985

	Conversion Factor tCO ₂ /toe	Electricity		Transportation	
		Mtoe	MtCO ₂	Mtoe	MtCO ₂
FRG					
Coal	3.556	44.18	157.11		
Oil	2.707	1.42	3.84	42.08	113.93
Gas	2.140	4.22	9.03		
Total		49.82	169.99	42.08	113.93
USA					
Coal	3.556	357.37	1270.86		
Oil	2.707	25.12	68.01	445.75	1206.85
Gas	2.140	72.66	155.49		
Total		455.15	1494.36	445.75	1206.85

Source: (OECD/IEA 88).

20% reduction in air pollution for the U.S. was \$17 billion in 1978 prices (Freeman 79). Assuming linearity, this implies total damages of \$85 billion per year. A more recent study for the EPA suggests that the annual benefits might be greater than Freeman's figure, but it is difficult to derive an annualized figure from the method used (Mathtech 83); see also (OECD 84).

For Germany, one study estimates the annual total costs of respiratory diseases (assuming 29% to 50% are due to air pollution) to range from DM 2.3 to DM 5.8 billion (Wicke 86). Another estimate ranges from 6.8 to 27.2 billion DM/yr (Grupp 86). Still another study yields damage estimates ranging from 1.6 to 40.3 billion DM/yr (Hohmeyer 88), based on figures given by (Euler 84).

Air pollution also causes accelerated corrosion and weathering of materials. For instance surface corrosion rates of zinc are five times higher in polluted areas than in clean-air regions, causing slow but significant damages to steel construction (Isecke 86). The need to enhance surface protection (of cables, for instance) increases costs in ways that are difficult to specify precisely. Further sources of damages include corrosion of natural stone used in buildings and outdoor works of art such as sculptures (The Statue of Liberty and Cologne Cathedral, for example).

Freeman estimated that benefits of a 20% reduction of air pollution for the U.S. to be about \$0.9 billion in the field of materials; this implies total damages (assuming linearity) of \$4.5 billion. In earlier times, when coal-burning was more widespread Parrish found the total annual damages to be some \$6 billion a year (Parrish 63).

For Germany annual damages from corrosion have been estimated by various people. One study set the figure at about 2.3 billion DM (Wicke 86), based on updated studies by (Heinz 86). Grupp estimated annual costs of 3.2–4.2 billion DM (not including damages to cultural property and monuments, in contrast to additional window cleaning). Hohmeyer (1988) estimated the flow of damages from corrosion to be in the range of 2.3 to 4.2 billion DM/yr.

Air pollution also causes vegetation damages by acidic deposition of photochemical oxidants (“acid rain”). Freeman estimated benefits of a 20% air pollution reduction in terms of reduced damage to vegetation in the range of \$0.2–\$2.4 billion/yr (Freeman 79). Total damages would, of course, be 5 × greater. Crocker arrived at more narrowly bounded figures of \$3.1 billion/yr, including forest losses \$1.75 billion, agricultural losses of \$1 billion and aquatic ecosystem damages of \$0.3 billion (Crocker 79).

In the FRG, the most visible and severe damages to vegetation are found in the forests. One study evaluated the forest decline as an annual loss of value of 5.5 to 8.8 billion DM (Ewers *et al.* 86). Hohmeyer estimated additional damages to agricultural crops of 1 billion DM/yr (Hohmeyer 88). No damages to wild flora and fauna were taken into account.

Table VI and VII summarize the numbers chosen by Freeman and Hohmeyer (Freeman 79; Hohmeyer 88). Note that both authors assume large confidence intervals which should be interpreted as the possible range of values of minimum cost figures. The figures are underestimates, to the extent that some important components are still omitted, for example loss of quality of life. Using Tables VI and VII we can relate damages to specific types of fuel consumption. Freeman’s damage figures are based on benefits of pollution control by 20% reduction, more than ten years ago. However, as shown in Figure 5 the level and composition of emissions in the U.S. hasn’t changed dramatically since the 70’s.

For the purposes of this study, the decline in emission is assumed to be

Table VI. Benefits of an air pollution reduction of 20% (billion 1978 US\$/yr)

Damages to US	Range	Best Point	Percent
Health	3.1–39.3	16.8	80%
Soil	0.5–5.0	2.0	9%
Vegetation	0.2–2.4	0.7	3%
Materials	0.5–1.4	0.9	4%
Other	0.1–8.9	1.0	4%
Total	4.4–57.0	21.4 (\$21.4: –80%, +166%)	100%

Source: (Freeman 79).

Table VII. Estimated damages from air pollution by type of damage (billion DM/year)

Damages to FRG	Range	Center	Percent
Health	1.6–40.3	21.0	65%
Materials	2.3–4.2	3.2	10%
Vegetation	6.5–9.5	8.2	25%
Total	10.4–54.0	32.2	100%

Source: (Hohmeyer 88).

compensated roughly by the increase in prices. For the U.S. transport sector, we impute benefits of $0.275 * \$107 = \29.4 billion/yr associated with 1315 Mt-CO₂/yr emissions, which yields a net benefit of \$22 per ton of CO₂ eliminated by consuming less petroleum based fuels. Similarly, for the electricity sector, we obtain $0.334 * \$107 = \35.7 billion/yr. Dividing by related emissions of 1532.54 Mt-CO₂ implies a benefit of \$23 per ton of CO₂ eliminated, mainly by consuming less coal.

Based on similar calculations for the FRG, air pollution damage from mobile sources amounts to 103 DM (roughly \$57 at current exchange rates) per ton of CO₂ eliminated. For electricity generation, we obtain 51 DM (\$28) per ton of CO₂ eliminated. The fact that the German figures are higher than the U.S. figures is understandable, in view of the fact that damages are a function of exposure, and Germany is a more densely populated country. It should be emphasized, again, that these figures are not precise. In fact, the range of uncertainty on the upper side is of the same order as the number, while on the down side it is of the order of half the number.

8. Conclusions for Policy

It is not our purpose to try to establish an optimal policy instrument to induce energy conservation, nor even an optimal tax policy (given the economists usual prejudice in favor of taxes over regulation). In this context, it seems to us that optimality is an inappropriate goal. Not only is it essentially impossible to achieve, but the time and effort expended on the debate is a distraction from the real issue. Current policies were not optimized from an economic perspective, nor is the political system conducive to optimization. But it is not difficult to assemble persuasive arguments that *almost any* tax on extractive resources — especially those whose intensive use results in large, unpaid social and environmental costs — is preferable to the present system of taxing labor (either directly or in the guise of value-added), assets (capital) and consumption.

Taxation is unquestionably an effective method of modifying social behavior, from consumption to investment, as economists have long recognized.⁹ A general principle of tax policy should be to reduce or eliminate taxes on desirable behavior (such as personal savings or capital investment) and to increase taxes on undesirable consumption items such as cigarettes and alcoholic beverages. With respect to the consumption of energy (mostly from fossil fuels) and other environment-destroying toxic substances, however, tax policy in the U.S. is contradictory. Most countries set very high taxes on all forms of energy use, especially automotive gasoline. (Despite high taxes, the energy and automotive industries of Europe and Japan remain healthy). The U.S. does not tax energy or toxic materials use, despite an obvious need for more government revenue, out of political concern for the jobs that might be lost.

With respect to revenues, the assumption made by many conservatives is that revenues siphoned out of the private sector by *any* tax will be used totally unproductively by the government. Liberals, on the other hand, seem to assume just the opposite. In the present context, the liberal assumption would be that revenues from a carbon tax (or a sulfur tax) would be automatically used to compensate for the environmental damages caused by the buildup of CO₂ or SO_x. The truth is certainly not so simple, either way. But, disregarding the use of the tax revenues, *It would still be socially beneficial to impose such a tax as long as its net cost to energy users (i.e. society) is less than the social benefits — reduced environmental damages — of lower fossil fuel combustion.*

For each tonne of CO₂ eliminated, we estimate these benefits as follows: \$30 for natural gas (based on greenhouse effects of CO₂ alone, ignoring likely methane leakage which would probably justify a much higher tax), \$52/ton for petroleum and \$53/ton for coal in the U.S. Two fifths of the benefits in the case of petroleum and coal are attributable to associated reductions in other pollutants, notably NO_x, SO_x, particulates and unburned hydrocarbons. For the Federal Republic of Germany (FRG) the petroleum and coal figures are, respectively, \$87 and \$58.¹⁰ In the case of the FRG, health-related benefits of reduced NO_x and SO_x outweigh climate benefits.

To obtain this assumed benefit, consumption must be reduced by raising the effective price to consumers. The reduction in consumption resulting from a rise in price (or tax) depends on the price elasticity of demand. (The less elastic the demand, the greater the reduction in consumption per increment of added price). Nobody knows exactly what the price elasticity of demand for petroleum or coal is, but a number that “seems more or less right” to most economists is -0.3 . That is, a fuel price increase of 1% would result in a drop of 0.3% in demand for that form of fuel. Roughly, a 100% increase in the price of fuel would result in a 30% cut in use.

Coal costs about \$50/ton. It is about 80% carbon. To get a reduction of

1 ton of CO₂ (worth more than \$30 in terms of climatic benefits and more than \$20 in terms of air quality benefits) consumption of 0.273 tons of coal-carbon must be eliminated. This corresponds to 0.34 tons of coal not burned. Assuming a price elasticity of -0.3 , this is the reduction that would be achievable by increasing the price to consumers by 100%. In short, any tax that doubles the price of coal (i.e. \$50/tonne) would save over \$50 in environmental costs to the society as a whole, while also producing about \$33 in net revenues for the government — to use productively, or not. In the U.S., based on a pre-tax consumption of 500 million tonnes per annum, such a tax would cut consumption by 170 million tonnes and generate \$16.5 billion in revenues.

Of course, there would also be both short term adjustment costs (e.g. shifts to natural gas or untaxed fuels) and long term “drag” costs on the economy. The effect of the higher coal price would be passed through, for instance, to the steel industry and to the consumers of steel products; similarly it would affect the cost of electricity and all users of electricity. The end result would be a structural shift further away from energy-intensive industries and towards energy-conservation services and non-fossil sources of energy (such as solar electricity).

Yohe and his colleagues investigated the potential drag effects of a 100% carbon tax (phased in over the 20 years from 2000–2020) or an equivalent consumption restriction (Yohe *et al.* 89). Their assumptions are based on work suggesting that the sudden increases in petroleum prices that occurred in 1973 and 1979 caused drops in productivity growth (Olson 88). It is assumed that any tax on fossil fuels would have the same depressing effect. Assuming productivity drops ranging from 0.05% p.a. to 0.7% p.a. for a carbon tax amounting to 100% of current prices, cumulative lost productivity of \$75 to \$550 billion would be experienced up to the year 2010.

This approach is not very convincing, for two reasons. First, the essence of the “oil shock” was that money was taken from western producers and consumers and deposited in bank accounts belonging to (predominantly) OPEC members. They, in turn, increased their spending for consumer goods, military goods and long-term infra-structure projects, none of which contributed to increased productivity in the west. On the other hand, the impact of a tax collected within the western economies would, of course, depend on how the money was spent. However, if the new tax were offset dollar-for dollar by a reduction in other taxes, there is no reason to expect an automatic depressing effect on the economy. Second, we note that the observed productivity drop since 1973–74 has many other possible explanations, of which the most widely accepted seems to be that it reflects the inevitable “catchup” of the U.S. by Europe and Japan.¹¹

In summary, we think a carbon tax (proportional to carbon content of different fuel types) and/or a sulfur tax (proportional to the sulfur content

of the airborne or water-borne emissions) would be consistent with the so-called “polluter-pays” principle and would avoid some of the administrative burden associated with regulation. In order to avoid severe market disturbances, such a tax should be slowly phased in (over one or two decades), in the context of international cooperation. We strongly suggest that the (rather high) *tax revenues could (and should) be used to reduce other forms of taxes*. As we have noted, such taxes now fall mainly on labor, capital and trade, which are economic activities that should be encouraged — insofar as they do not involve the use of fossil fuels or toxic materials — rather than discouraged.

It is interesting to compare our final results with those of Nordhaus (Nordhaus 89), in terms of marginal shadow costs of abatement. The Nordhaus analysis assumed that the cost of abatement rises monotonically with the degree of abatement, i.e. percentage of CO₂-equivalent reduction. The major policy conclusions of his analysis were as follows:

- CFC phase-out is the lowest cost option. Total elimination would cut the CO₂-equivalent emissions by 14%. A high gasoline tax would be the next most effective policy option.
- Reforestation is not a cost-effective option (at marginal costs assumed to be \$100/ton of CO₂ removed).
- The shadow costs of CO₂ reduction by a global carbon tax would exceed the benefits except at low levels of abatement. It is not an efficient policy.

For the case of high costs associated with greenhouse warming (\$37/ton of carbon dioxide or equivalent RIG) the optimal policy would yield a 28% reduction in CO₂-equivalent emissions and 94% reduction of CFC use.

We have arrived at significantly different conclusions, partly because we disagree with the underlying cost assumptions, and partly because we have tried to take into account ancillary benefits of policies that would reduce other pollution-related damages.

In our view, as stated above, the most cost-effective policy is likely to be energy conservation combined with substitution of natural gas for other fuels to the extent dictated by direct cost savings. This would reduce energy consumption by 20% or so and carbon dioxide emissions by about 25% compared to present levels, though the overall impact on RIG emissions would be more like 12%. The most effective policy to bring about this degree of conservation would be a carbon tax on all fossil fuels supplemented by a sulfur emissions tax on sulfur-containing fuels.

The virtual elimination of CFC's, as suggested by Nordhaus, would probably be the next most cost-effective policy, with a further 14% reduction in RIG's. Further abatement beyond that would be achievable by some combination of further energy conservation, fuel substitution, and introduction of non-polluting solar (or other) energy sources, combined with exten-

sive reforestation. We are unable to evaluate the relative costs and benefits of these policies in more detail at this time, however.

Appendix A: Forests

A popular proposal is to sequester excess atmospheric carbon-dioxide by means of photosynthesis, using large plantations of fast-growing trees as a sink. To halt the present annual atmospheric increase of carbon (2.9 Gt of C per year) by massive afforestation, the required land area in temperate climate regions would be about 465 million hectares (Mha) (Sedjo & Solomon 90). This is equal to 1.5 times the present forest area of the U.S. or 15% of the global forest area. The land requirement is large but not totally out of the question. For instance, it is biologically feasible to double the biomass density (Sedjo & Solomon 90; Ranney *et al.* 87; Myers & Goreau 90)

Planting costs are estimated to be \$230–\$1000/ha (averaging \$400/ha). In the U.S. suitable land costs roughly \$400–\$1000/ha, although in some regions it would be lower. Land costs represent real opportunity costs, since land used as carbon sink cannot be used for crops or pasture. The prospects of finding a large area of suitable land in the western countries appears slim. However, it is estimated that there is at least 500 Mha of degraded or deforested tropical land (Houghton & Woodwell 89). Degraded land costs much less; in Indonesia total costs of \$400 per hectare are indicated (Sedjo & Solomon 90); in India, costs are even lower (Myers 89).

Total costs of \$372 billion for new plantations in temperate zones and \$186 billion in the tropics, would result in marginal costs of reducing CO₂ of \$35/ton and \$17/ton respectively (Sedjo & Solomon 90). These figures are in sharp contrast to the pessimistic \$100/ton estimated by Nordhaus (1989).

Appendix B: Reducing CFC's

Chlorofluorocarbons (CFC's) are chemically inert compounds, used as solvents, propellants and refrigerants. They are not readily degraded by chemical reactions in the lower atmosphere. As a consequence, CFC's gradually diffuse into the stratosphere where they are effective absorbers of long-wave (IR) thermal radiation. Because CFC's are about 20,000 times more efficient than CO₂ as IR absorbers, the two most common CFC's, CFC-11 and CFC-12, alone contribute about 14% of all RIG emissions.

The replacement of CFC's is relatively cost efficient because of the small quantities involved. The propellant and solvent uses are fairly easy to find substitutes for. The most difficult use to dispense with is refrigeration and air-conditioning. However, there are a variety of technically feasible sub-

stitutes for CFC's with shorter atmospheric lifetimes and/or less absorption strength in the critical frequency band. Despite some disadvantages, such as greater flammability and reduced efficiency as refrigerants, it is reasonable to assume substitution costs of the order of \$5/kg, or about \$0.25/ton of CO₂ equivalent eliminated (Pool 88).

Problems will arise in developing countries. China, for instance, is planning to raise refrigerator production by the year 2000, and to boost CFC production to levels tenfold higher than those in the USA today; China did not sign the Montreal Protocol (Miller 89).

A non-trivial incidental benefit of CFC reduction would be a slower rate of depletion of the stratospheric ozone layer. We have no quantitative estimate of monetary worth of this, however.

Notes

¹ For details see: (Schneider 89, 89a, 89b; Schneider & Rosenberg 90; Bolin *et al.* 86; USEPA 88; Mintzer 87).

² Gross National Income (GNI) differs from GNP by excluding indirect business taxes and capital set aside to replace depreciation.

³ See, for instance, the Brundtland Report (Brundtland 87, pp. 291–294, 300); also (Renner 89, pp. 141–144); (Myers 89).

⁴ The classic example of the “lock-in” phenomenon is the QWERTY typewriter keyboard, which is known to be inefficient but which is so well established that it seems unchangeable (David 85). Another example might be the persistent use of the so-called “English system of weights and measures.” For a theoretical discussion of positive returns to scale and self-reinforcing mechanisms in economics see (Arthur 83, 88; Arthur *et al.* 87a).

⁵ For an early review paper see (Manne, Richels & Weyant 79). Also (Manne 81; Edmonds & Reilly 85, 85a; Manne & Richels 90).

⁶ Many projects are evaluated in terms of *payback time* rather than *return*. The two concepts are closely related. A project with a payback time of 1 year corresponds to 100% return on investment. A project that pays for itself in 6 months has an annual return of 200%, and so on.

⁷ It should be noted that the Mellon study was thoroughly criticized by a group at MIT, at the request of the Department of Energy (Berndt *et al.* 8a). The critique was extensive and detailed, and a number of significant substantive and methodological criticisms were offered. One criticism was directed at the study's implication that a “least cost” solution would be achieved automatically if the economy were truly competitive. The authors of the critique asserted that competition does not necessarily yield an optimal result and that regulation might be more effective. The critique also noted that some of the projected savings were “imposed” on the study, rather than being derived endogenously. The examples cited in this regard included projected savings by the use of variable-speed electric motors, co-generation of electric power and industrial process heat, and dieselization of the bus fleet.

In retrospect, the benefits of variable speed motors were probably exaggerated somewhat. However, in defense of the study, it should be noted that there is no way technological shifts such as the ones noted could be generated endogenously by any economic model. Moreover, a large number of other specific but minor opportunities for saving energy via the use of available technology were necessarily overlooked, simply because the authors had limited time and resources available to them. Thus, it is more likely that the extent of the conservation opportunities were underestimated than conversely.

⁸ Losses of quality of life are still not monetarized, hence not included in any of the estimates below.

⁹ The major argument of the “supply-siders” in the early Reagan years was that excessive taxes on income would discourage productive effort, whence tax-cuts would actually generate new entrepreneurial activity. The long-standing argument for eliminating capital gains taxes is based on similar notions. Virtually all economists would agree that income or capital gains tax cuts are stimulative. The major argument is whether the revenue gains for the government would exceed the losses, which is a very different issue.

¹⁰ To convert these benefit numbers to \$ per ton of actual fuel, remember that each ton of carbon dioxide emitted corresponds to 0.273 tons of carbon in the fuel, so the benefit is in proportion.

¹¹ In fact, an OECD Seminar on the “Apparent Productivity Paradox” held in June 1989 and attended by a number of the world’s top economists, considered in detail a number of the extant theories of the decline in productivity growth since 1972–73. The rise in energy prices as a causal factor was not even discussed. This does not mean that energy prices played no role, but it appears that most economists who attended no longer believe that role to have been crucial.

References

- Arthur, W. Brian (1988), *Competing Technologies and Lock-In by Historical Small Events: The Dynamics of Allocation under Increasing Returns*, Research Paper (43), Committee for Economic Policy Research, Stanford University, Palo Alto CA.
- Arthur, W. Brian (1988), ‘Competing Technologies: an Overview’, in Dosi *et al.* (eds.), *Technical Change and Economic Theory*: 590–607, Pinter Publishers, London.
- Arthur, W. Brian *et al.* (1987), ‘Path-dependent Processes and the Emergence of Macrostructure’, *European Journal of Operations Research* **30**, 294–303.
- Arthur, W. Brian, Yu M. Ermoliev, and Yu M. Kaniovski (1987), *Non-Linear Urn Processes: Asymptotic Behavior and Applications*, Working Paper (WP-87-85), International Institute for Applied Systems Analysis, Laxenburg, Austria.
- Batie, S. S. and H. H. Shugard (1990), ‘The Biological Consequences of Climate Changes: An Ecological and Economic Assessment’, in Rosenberg (ed.), *Greenhouse Warming: Abatement and Adaptation*: 121–132, Resources for the Future, Washington DC.
- Benkovitz, C. (1982), ‘Compilation of an Inventory of Anthropogenic Emissions in the United States and Canada’, *Atmospheric Environment* **16**(6), 1551–1563.
- Berndt, E. R., M. Manove, and D. O. Wood (1981), *A Review of the Energy Productivity Center’s “Least-Cost Energy Strategy” Study*, Massachusetts Institute of Technology, Cambridge, MA.
- Bolin, Burt, Bo R. Döös, and Jill Jaeger (1986), *The Greenhouse Effect, Climatic Changes and Ecosystems*, John Wiley and Sons, Chichester.
- Brundtland, G. H. (ed.) (1987), *Our Common Future*, Oxford University Press, New York, (Report of the WCED.)
- Crocker, T. (1979), *Experiments in the Economics of Air Pollution Epidemiology*, Report (66/5-79001a), United States Environmental Protection Agency, Washington DC.
- Crosson, Peter R. (1990), ‘Climate Change: Problems of Limits and Policy Responses’, in Rosenberg (ed.), *Greenhouse Warming: Abatement and Adaptation*, 69–82, Resources for the Future, Washington DC.
- David, Paul A. (1985), ‘Clio and the Economics of QWERTY’, *American Economic Review* (Papers and Proceedings) **75**, 332–337.
- D’Errico, Emilio, Pierluigi Martini, and Pietro Tarquini (1984), *Interventi di risparmio energetico nell’industria*, Report, ENEA, Italy.

- Edmonds, J. and J. M. Reilly (1985), *Global Energy – Assessing the Future*, Oxford University Press, New York.
- Edmonds, J. and J. M. Reilly (1985d), 'Future Global Energy and Carbon Dioxide Emissions', in Trabalka (ed.), *Atmospheric Carbon Dioxide and the Global Carbon Cycle* (DOE/ER-039), National Technical Information Service, Springfield VA.
- Euler, H. (1984), *Umweltverträglichkeit von Energiekonzepten, Planungsgrundlagen für die Erstellung von umweltorientierten örtlichen und regionalen Energieversorgungskonzepten*, Bonn, FRG.
- Ewers, H. J. et al. (1986), *Methodische Probleme der monetären Bewertung eines komplexen Umweltschadens – Das Beispiel des Waldschadens in der Bundesrepublik Deutschland*, Report (FB 101 03086), Bundesrepublik Deutschland Umweltbundesamt, Berlin.
- Frederick, K. D. and Peter H. Gleick (1990), 'Water Resources and Climate Change', in Rosenberg (ed.), *Greenhouse Warming: Abatement and Adaptation*, 133–146, Resources for the Future, Washington DC.
- Freeman, A. Myrick III (1979), *The Benefits of Environmental Improvement: Theory and Practice*, Johns Hopkins University Press, Baltimore.
- Freeman, Christopher (1982), *The Economics of Industrial Innovation*, MIT Press, Cambridge MA, 2nd edition.
- Gaines, L. et al. (1979), *TOSCA: The Total Social Cost of Coal and Nuclear Power*, Ballinger Publishing Company, Cambridge, MA.
- Glantz, M. H. (1988), 'Societal Response to Regional Climate Change: Forecasting by Analogy', in Glantz (ed.), *Workshop*, Boulder CO.
- Grupp, H. G. (1986), 'Die sozialen Kosten des Verkehrs', *Verkehr und Technik*, Heft 9, pp. 359–366 and Heft 10, pp. 403–407.
- Hausman, Jerry A. (1979), 'Individual Discount Rates and the Purchase and Utilization of Energy-using Durables', *Bell Journal of Economics* **10**(1), 33–54.
- Heinz, I (1986), *Zur ökonomischen Bewertung von Materialschäden durch Luftverschmutzung*, Umweltamt.
- Hekstra, G. P. (1990), 'Sea-Level Rise: Regional Consequences and Responses', in Rosenberg (ed.), *Greenhouse Warming: Abatement and Adaptation*, 53–68, Resources for the Future, Washington DC.
- Hohmeyer, Olav (1988), *Social Costs of Energy Consumption*, Springer-Verlag, Heidelberg FRG.
- Houghton, Richard A. and George M. Woodwell (1989), 'Global Climactic Change', *Scientific American* **260**(4), 36–44.
- Isecke, B. (1986), *Einfluss von Luftverunreinigungen auf das Korrosionsverhalten verschiedener Materialien*, Umweltamt.
- Jacobson, J. (1989), *Abandoning Homeland in State of the World 1989*, World Watch Institute, Washington DC.
- Lave, Lester and E. Seskin (1971), *Air Pollution and Human Health*, Johns Hopkins University Press, Baltimore.
- Lovins, Amory B. et al. (1981), *Least-Cost Energy: Solving the CO₂ Problem*, Brickhouse Publication Co., Andover MA.
- Manne, Alan S. (1981), *ETA-MACRO: A User's Guide*, Report (EA-1724), Electric Power Research Institute, Palo Alto CA.
- Manne, Alan S. and Richard G. Richels (1990), 'Global CO₂ Emission Reductions: The Impacts Rising Energy Costs', in Ferrari, Tester and Woods (eds.), *Energy and the Environment in the 21st Century*, MIT Press, Cambridge MA.
- Manne, Alan S., Richard G. Richels, and J. P. Weyant (1979), 'Energy Policy Modeling: A Survey', Operations Research Society of America', *ORSA J.*
- Miller, J. (1989), 'Chinese Bring Chill to Backers of Ozone Protocol', *New Scientist* **11**, 28.

- Mintzer, Irving M. (1987), *A Matter of Degree: The Potential for Controlling the Greenhouse Effect*, Research Report (5), World Resources Institute, Washington DC, April 1987.
- Myers, N. (1989), 'The Environmental Basis of Sustainable Development', in Scramm and Warford (eds.), *Environmental Management and Economic Development*, Johns Hopkins University Press, Baltimore.
- Nelson, Kenneth E. (1989), 'Are There Any Energy Savings Left?', *Chemical Processing*.
- Nordhaus, William D. (1973), 'The Allocation of Energy Resources', *Brookings Papers on Economic Activity* 3.
- Nordhaus, William D. (1975), 'The Demand for Energy: An International Perspective', in Nordhaus (ed.), *Workshop on Energy Demand*, International Institute for Applied Systems Analysis, Laxenburg, Austria.
- Nordhaus, William D. (1989), 'The Economics of the Greenhouse Effect', in *International Energy Workshop*, International Institute for Applied Systems Analysis, Laxenburg, Austria.
- Nordhaus, William D. (1990), 'Economic Policy in the Face of Global Warming', in Ferrari, Tester, and Woods (eds.), *Energy and the Environment in the 21st Century*, MIT Press, Cambridge MA.
- Nordhaus, William D. and Gary Yohe, (1983), 'Future Carbon Dioxide Emissions from Fossil Fuels', in *Changing Climate*, National Academy Press. (National Research Council-National Academy of Sciences.)
- Olson, Mancur (1988), 'The Productivity Slowdown, the Oil Shocks, and the Real Cycle', *Journal of Economic Perspectives* 2 (4), 43–69.
- Parrish, E. M. (1963), *Effects of Air Pollution Property Damages and Visibility*, Pennsylvania Air Pollution Control Institute, Harrisbury PA.
- Pool, R. (1988), 'The Elusive Replacements for CFCs', *Science* 242, 666–668.
- Ramanathan, V. (1988), 'The Radiative and Climate Consequences of the Changing Atmospheric Composition of Trace Gases', in Rowland and Isaksen (eds.), *The Changing Atmosphere*, John Wiley, USA.
- Ranney, J. W., L. L. Wright, and P. A. Layton (1987), 'Hardwood Energy Crops: The Technology of Intensive Culture', *Journal of Forestry* 85, 17–28.
- Renner, M. (1989), *Enhancing Global Security in State of the World 1989*, World Watch Institute, New York.
- Robin, G. de Q. (1986), 'Changing the Sea Level', in Bolin *et al.* (eds.), *The Greenhouse Effect, Climatic Change and Ecosystems*, John Wiley and Sons, Chichester UK.
- Robinson, John B. (1987), 'An Embarrassment of Riches: Canada's Energy Supply Resources', *Energy* 12 (5), 379–402.
- Robinson, John B. (1987), 'Insurmountable Opportunities? Canada's Energy Efficiency Resources', *Energy* 12 (5), 403–417.
- Ross, M. *et al.* (1975), *Effective Use of Energy: A Physics Perspective*, Technical Report, American Physical Society (Summer Study on Technical Aspects of Efficient Energy Utilization).
- Sant, R. W. (1979), *The Least-Cost Energy Strategy: Minimizing Consumer Costs Through Competition*, Report (55), Mellon Institute Energy Productivity Center, VA.
- Schipper, Lee (1989), 'Energy Efficiency in an Era of Apparent Energy Stability: Progress, Plateau, or Passé?', in *International Energy Workshop*, International Institute for Applied Systems Analysis, Laxenburg, Austria.
- Schneider, Stephen H. (1989), 'The Changing Climate', *Scientific American* 216 (3), 38–47.
- Schneider, Stephen H. (1989a), 'The Greenhouse Effect: Sciences and Policy', *Science* 243.
- Schneider, Stephen H. (1989b), *Global Warming: Are We Entering the Greenhouse Century?*, Sierra Club Books, San Francisco.
- Schneider, Stephen H. and Norman J. Rosenberg (1990), 'The Greenhouse Effect: Its

- Causes, Possible Impacts and Associated Uncertainties', in Rosenberg (ed.), *Greenhouse Warming: Abatement and Adaptation*, 7–34, Resources for the Future, Washington DC.
- Sedjo, R. A. and Allen M. Solomon (1990), 'Climate and Forests', in Rosenberg (ed.), *Greenhouse Warming: Abatement and Adaptation*, 105–120, Resources for the Future, Washington DC.
- Socolow, R. H. (1975), 'Efficient Use of Energy', *Physics Today*, 23–33.
- Wicke, L. (1986), *Die ökologischen Milliarden*, Kösel-Verlag, Munich.
- Wigley, T. M. L., *Climate Monitor* **16**, 14–28.
- Wilson, Edward O. (1989), 'Threats to Bio-diversity', *Scientific American* **261** (3), 66–66.
- Yohe, G., D. Howardh, and P. Nikiopoulus (1989), *On the Ability of Carbon Taxes to Fend Off the Greenhouse Warming*, Department of Economics, University of Connecticut, Middletown CT.
- Mathtech Inc. (1983), *Benefits and Net Benefit Analysis of Alternative National Ambient Air Quality Standards for Particulate Matter*, Mathtech Inc. (Prepared for the Economic Analysis Branch, Office of Air Quality Planning and Standards, U.S. Environmental Planning Agency (5 Vols)).
- National Acid Precipitation Assessment Program (1990), *Integrated Assessment; Questions 1 and 2*, draft final report (1), National Acid Precipitation Assessment Program, Washington DC.
- Organization for Economic Cooperation and Development (1984), *Background Papers*, International Conference on Environment and Economics, Organization for Economic Cooperation and Development, Paris.
- Organization for Economic Cooperation and Development (1989), *Environmental Data Compendium 1989*, Organization for Economic Cooperation and Development, Paris.
- Organization for Economic Cooperation and Development/International Energy Agency (1988), *Energy Balances 1985–1986*, Organization for Economic Cooperation and Development/International Energy Agency, Paris.
- United States Environmental Protection Agency (1988), *The Potential Effects of Global Climate Change on the United States*, Draft Report to Congress, United States Environmental Protection Agency, Washington DC.
- United States Environmental Protection Agency (1984), *Policy Options for Stabilizing Global Climate*, Draft Report to Congress, United States Environmental Protection Agency, Washington DC.
- World Resources Institute (1989), *World Resources Yearbook 1988–89*, World Resources Institute, Washington DC.